EUROPEAN PURSE SEINE CPUE STANDARDIZATION: METHODOLOGY AND FRAMEWORK FOR THE YFT STOCK ASSESSMENT

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SUMMARY

Purse seine CPUE standardization is thought on a combination of fishing mode and commercial size categories of species basis, i.e., large fish in free schools (FSC) sets on one side and small fish under floating objects associated sets on the other side. However, while FSC sets are randomly encountered, FOB sets can either be randomly encountered, e.g., foreign drifting fish aggregating devices (dFADs) or natural log not instrumented, or not randomly encountered, i.e., vessels have access to buoys and/or echosounder data equipping the dFAD. The non-randomness of encounter leads to different statistical approaches and different impact on effort creep. On one hand, the standardization approach using an extension of the Delta-lognormal GLMM to three components, *i.e., the product of the number of schools detected (summing positive and null sets) (number of schools) by spatio*temporal strata, the proportion of positive sets with the species/category of interest and the catch per positive set with it (school size), is appropriate to randomly encountered schools. We propose to apply this methodology to FSC sets as well as to FOB sets randomly encountered. On the other hand, for FOB sets not randomly encountered, we propose to use, as classical approaches, the product of the third component, i.e., school size, by a fishing efficiency rate per set calculated with a methodology quantifying the increase in fishing efficiency due to the use of FOB equipped with echosounders (Wain et al. 2020). This framework would allow to homogenized standardization of CPUE based on fisheries-dependant data and provide several time series, i.e., on randomly encountered FSC (> and < 10kg) and FOB sets separately and on not randomly encountered FOB sets, here of EU purse seine fleet catches per unit effort (CPUE) of yellowfin tuna (YFT) from the Indian Ocean.

KEYWORDS: Abundance; CPUE standardization; Purse seiner; Spatial variations; Logbooks; Yellowfin tuna

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1. Introduction

The goal of this paper is to propose a development of yellowfin tuna (YFT, *Thunnus albacares*) indices of abundance derived from reported catches of EU tropical tuna purse seiners operating in the Indian Ocean. This work thus aimed at developing standardised Catch per unit effort (CPUE) time series to be provided to IOCT as an input for the stock assessment of YFT.

We propose to compare EU standardized CPUE:

- Using the 3 components methodology, adapted for randomly encountered schools, presented in Guéry *et al.* (IOTC-2019-WPM10-14) and attached below, on:
 - Sets with adult YFT (> 10kg) in FSC randomly encountered since 1991
 - Sets with juvenile YFT (< 10kg) in FSC randomly encountered since 1991
 - Sets with juvenile YFT (< 10kg) under FOB randomly encountered since 2010
- Using the traditional approach for comparison (1 component approach)
 - \circ on all sets with juvenile YFT done under FOB
- Using the traditional approach (1 component approach) multiplied by a fishing efficiency rate per set following the methodology published in Wain *et al.* (2020) attached below:
 - o Sets with juvenile YFT under FOB not randomly encountered

ACCOUNTING FOR FISHING DAYS WITHOUT SET, FISHING CONCENTRATION AND PIRACY IN THE CPUE STANDARDISATION OF YELLOWFIN TUNA IN FREE SCHOOLS FOR THE EU PURSE SEINE FLEET OPERATING IN THE INDIAN OCEAN DURING THE 1991-2017 PERIOD

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SUMMARY

The time series of EU purse seine fleet catches per unit effort (CPUE) of yellowfin tuna (YFT) from the Indian Ocean were standardized using an extension of the Delta-lognormal GLMM to three components. The aim was to depict the trend in abundance for adult YFT observed in free schools (FSC). The originality of this work relied on the inclusion of i) null sets, considered as presence of YFT FSC, ii) fishing days without set, considered as absence of FSC, iii) EU fishing agreement in the exclusive economic zones driving EU purse seine fleet presence in these areas, iv) time spent by centroid cell by boat by day to constrain detectability, v) the Gulland's index of fishing effort concentration to measure the extent to which a fleet has concentrated its fishing effort in areas with higher than average catch rates and, vi) piracy as a presence absence variable. Standardized CPUE for FSC was thus defined as the product of the number of set (positive and null) by spatio-temporal strata, the proportion of sets with large YFT (>10 kg) and the catch per large YFT set. To detect strata without sets, all activities recorded in captain logbooks were used for the period 1991-2017. This new standardization approach, therefore, represents a significant advance over previous efforts, though there are a number of avenues for future progress.

KEYWORDS: Abundance; CPUE standardization; Purse seiner; Spatial variations; Logbooks; Yellowfin tuna

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1. Introduction

The goal of this paper is to develop yellowfin tuna (YFT, *Thunnus albacares*) indices of abundance derived from reported catches of EU tropical tuna purse seiners operating in the Indian Ocean, as part of the EASME/EMFF/2016/008/SC14. This work thus aimed at developing standardised Catch per unit effort (CPUE) time series to be provided to IOCT as an input for the stock assessment of YFT. In this study we considered only sets of adult YFT (> 10 kg) on free schools (FSC) so as to capture the spawning-stock biomass of the population.

Interpreting changes over time in CPUE series as a trend in abundance has always been a major challenge for scientists working in stock assessment. In the case of the tropical tuna purse seiner fishery operating in the Indian Ocean, there are several factors affecting the CPUE-abundance relationship, such as the increase in individual fishing power of the vessel due to the implementation of new technologies (Gaertner and Pallares, 2002, Fonteneau *et al.*, 1999), the change in the fishing grounds, or the extensive development of fishing on drifting fish aggregative devices in the early 1990s (Hallier and Parajua, 1992), the Somali piracy issue that increased since the early 1990s within the Gulf of Aden and progressively expanded toward the east of the Indian Ocean (Chassot *et al.* 2010) and the high variability in fishing effort concentration due to anomalies in the distribution of tuna resources (Fonteneau *et al.* 2008, Potier *et al.* 2004). In addition, as tunas are usually spatially structured in schools and in clusters of schools, it is important to consider that any change in abundance may be influenced by the number (or density) of schools at sea, as well as by the size of individual schools (Laurec et Le Guen, 1977; Fréon and Misund, 1999). To account for the presence of a large number of zerocatch fishing days, the delta-lognormal method (Lo *et al.*, 1992) has commonly been used in a variety of fisheries, the specific index for a given year being the product of year average fitted values of the lognormal (for the positive CPUEs) and the binomial (for the proportion of days with catch) models.

With these considerations in mind, Katara et al. (2016, 2017) developed a delta-lognormal GLMM approach with two sub-models: a binomial GLMM that standardises the probability of a positive set, and a lognormal GLMM that standardises catch conditional on the set being positive. In this approach, a positive set is taken to be a set with catch of the target species, so null sets are implicitly taken to indicate the absence of fish. However, bearing in mind the fact that null sets generally correspond to the presence of a fish school that simply avoided the net and that the abundance of large YFT may also be related to the number of free schools detected per unit of search time, we used in this paper a new modelling approach adopted for the Atlantic YFT stock assessment (Guéry et al. 2019) based on a Delta-lognormal GLMM with 3 components: (i) a Poisson first component modelling the density of free schools based on the number of sets (positive and null) per unit of single-boat searching time, (ii) a binomial second component modelling the presence of large YFT in free school positive sets (i.e., sets with catch of any species; non-null sets), and (iii) a lognormal third component modelling the amount of large YFT per free school positive set with large YFT. Along with the commonly used covariates related to vessel characteristics and spatiotemporal variability, the originality of this work consists of the inclusion of i) null sets, considered as presence of YFT FSC, ii) fishing days without sets, considered as absence of FSC, iii) data on EU fishing agreements (or lack thereof) in exclusive economic zones of regional countries driving EU purse seine fleet presence in these areas, iv) search time spent by centroid cell by boat by day to constrain detectability, v) the Gulland's index (Gulland 1956, Fonteneau 1982, Gaertner 1984, Hoyle et al. 2017) of fishing effort concentration to measure the extent to which a fleet has concentrated its fishing effort in areas with higher than average catch rates and, vi) piracy as a presence/absence variable. Standardized CPUE for FSC was thus defined as the product of the number of sets (positive and null) per unit search time, the proportion of positive (non-null) sets containing large YFT (> 10 kg) and the catch per set of large YFT. To detect and include cells explored, but for which a set was not carried out, all activities recorded in the captain logbooks were used for the period 1991-2017.

2. Material and Methods

2.1. Conventional fishing data

Logbook data for the French and Spanish purse seine fleets targeting tropical tuna in the Indian Ocean from 1991 to 2017 were analysed to derive the standardised CPUEs. The logbook databases are managed by the Tuna Observatory (Ob7) and the IEO for the French and the Spanish fleets, respectively. The raw logbook data (Level 0) produced by the skippers were corrected in terms of total catch per set (to account for the difference between

reported catch at sea and landed catch) and species composition (based on port size sampling and the T3 methodology – see Pallarès and Hallier 1997) to generate the Level 1 logbook database used in this paper. The free-school sets (FSC) dataset, i.e. non-associated school sets and whales' sets, was used to derive CPUE for the adult fraction of the YFT stock, by selecting the commercial size categories 2 and 3 (> 10 kg). The analysis was restricted to:

- The period 1991-2017 for FSC sets to be able to have the larger coverage of days without set reported
- 2018 was removed from the analyses due to a change in fishing strategy where FSC were not targeted
- The area defined by all grid cells where YFT of category 2 and 3 were fished for at least 5 years over a period of no less than 15 years, to avoid areas that are not routinely fished
- High seas and all EEZs
- Vessels with fewer activities than the 5% of the left hand distribution based on the cumulative number of days per boat (all activities confounded) were removed
- Entire days with at least one activity with problematic operations were removed
- All sets per boat and day were aggregated and attributed to the centroid of these set activities
- Distances between successive sets null-FSC/next-FSC for a boat is not significantly different from all other combinations: no need of buffer avoiding to count the same school several times
- Total number of sets per day per boat was filtered and days with unrealistic data were removed

2.2. Modelling approach: Delta-lognormal GLMMs

As mentioned in the Introduction section, delta-lognormal GLMMs were developed including three sub-models: a Poisson GLMM that standardises the number of positive and negative sets, a binomial GLMM that takes into account the fraction of positive sets with large YFT and a lognormal LMM to describe the catch conditional to positive set. Available variables are detailed in **Table 1**.

Variable	Description
Fleet country	France; Spain
Numbat	Unique vessel identifier
Vessel storage capacity	In m ³
Gulland's index of fishing	Measure the extent to which a fleet has concentrated its fishing effort in areas with
effort concentration	higher than average catch rate (Gulland 1956, Fonteneau 1981, Gaertner 1984,
	Hoyle <i>et al.</i> 2017)
Number of sets on FOBs	Monthly resolution per grid cell
Number of positive set	Number of positive sets per boat per day per centroid
Year	Year at which the fishing set took place
Quarter	Quarter of years
Age of vessel	Year – Year of vessel service
Economic Exclusive Zone	Identifiers of EEZs and the offshore area
Fishing access	EU fishing agreement in the different EEZ. Binary variable by cell.
Piracy	Presence/absence of piracy per cell
Searching centroid	Single-boat searching time in hours calculated as (sun set time - sun rise time) -
	(number of set*median of setting time)

Table 1. Available variables for the calculation of CPUE and the development of the standardisation models

We performed the Poisson GLMM where the full model included the following fixed effects: fleet country, age of the vessel, number of sets on FOB, vessel storage capacity, year, quarter, Gulland index and piracy variable. The number of FOB sets per trip was included as a proxy for vessels' fishing strategy changes across time due to the increase of dFADs. The random structure of the model includes fishing access and a vessel unique identifier. The time spent by single-boat searching centroid by day was calculated as (sun set time – sun rise time) – (number of set*median of setting time) and was used as an offset.

Component 1:

 $num_sets_fsc \sim fleet \ country + age \ of \ the \ vessel + num_sets_fob + vessel \ storage \ capacity + year + quarter + gulland \ index + piracy + (1| numbat) + (1|eez:fishing_access) + offset \ (searching_centroid)$

The full model for the binomial GLMM and the lognormal LMM included the following fixed effects: fleet country, vessel storage capacity, year, quarter, Gulland index. The random structure of these models included a vessel unique identifier. The number of positive sets was used as an offset, as data were aggregated by boat, day and centroid cell.

Component 2:

 $yft_{pos} \sim fleet country + vessel storage capacity + year + quarter + gulland index + (1 | numbat) + offset(nb of positive sets)$

Component 3:

 $log_capture \sim fleet country + vessel storage capacity + year + quarter + gulland index + (1|numbat) + offset(nb of positive sets)$

GLMM tables and results are presented in appendices.

3. Results

3.1. FSC sets (1991-2017 period): Poisson GLMM (number of large-size YFT catch > 0 and = 0)

3.1.1. Diagnostics

Diagnostics are presented below (Figure 1, Figure 2, Erreur ! Source du renvoi introuvable. Figure 3).



box-plots of residuals by year

Figure 1. FSC sets - number of large-size YFT catch > 0 and = 0: box-plots of residuals by year

Histogram of residuals: Comp1 Poisson distribution



Figure 2. FSC sets – number of large-size YFT catch > 0 and = 0: histogram of residuals



Residuals vs. Fitted

Figure 3. FSC sets – number of large-size YFT catch > 0 and = 0: residuals versus fitted

3.1.2. Standardized time series

Standardized time series are presented by year-quarter (Figure 4) and by year (Figure 5). We used the package *arm v1.10-1* (*sim* and *fitted* functions) to simulate posterior distributions and calculate confidence intervals.

Component 1 : Poisson distribution



Figure 4. FSC sets – predicted number of large-size YFT catch > 0 and = 0: standardised time series by yearquarter (black) with 97.5% confidence intervals (grey) compared to nominal (red)



Component 1 : Poisson distribution

Figure 5. FSC sets – predicted number of large-size YFT catch > 0 and = 0: standardised time series by year (black) with 97.5% confidence intervals (grey) compared to nominal (red)

3.2. FSC sets (1991-2017 period): Binomial GLMM (fraction of positive set with large YFT)

3.2.1. Diagnostics

Diagnostics are presented below (Figure 6, Figure 7, Figure 8).





Figure 6. FSC sets - fraction of positive set with large YFT: box-plots of residuals by year



Histogram of residuals: Comp2 Binomial distribution

Figure 7. FSC sets - fraction of positive set with large YFT: histogram of residuals





Figure 8. FSC sets - fraction of positive set with large YFT: residuals versus fitted

3.2.2. Standardized time series

Standardized time series are presented by year-quarter (Figure 9) and by year (Figure 10). We used the package *arm v1.10-1* (*sim* and *fitted* functions) to simulate posterior distributions and calculate confidence intervals.



Component 2 : Binomial distribution

year_quarter

Figure 9. FSC sets – fraction of positive set with large YFT: standardised time series by year-quarter (black) with 97.5% confidence intervals (grey) compared to nominal (red)



Component 2 : Binomial distribution

Figure 10. FSC sets – fraction of positive set with large YFT: standardised time series by year (black) with 97.5% confidence intervals (grey) compared to nominal (red)

3.3. Log-Normal GLMM (catch per hour conditional to YFT catch > 0)

3.3.1. Diagnostics

Diagnostics are presented below (Figure 11, Figure 12, Figure 13).

box-plots of residuals by year



Figure 11. FSC sets – catch | catch > 0 with large YFT: box-plots of residuals by year



Histogram of residuals(this_model)

Figure 12. FSC sets – catch | catch > 0 with large YFT: histogram of residuals





Figure 13. FSC sets – catch | catch > 0 with large YFT: normal Q-Q plot

3.3.2. Standardized time series

Standardized time series are presented by year-quarter (Figure 14) and by year (Figure 15). We used the package *arm v1.10-1* (*sim* and *fitted* functions) to simulate posterior distributions and calculate confidence intervals.



Component 3 : Lognormal distribution

Figure 14. FSC sets – catch | catch > 0 with large YFT: standardised time series by year-quarter (black) with 97.5% confidence intervals (grey) compared to nominal (red)

Component 3 : Lognormal distribution



Figure 15. FSC sets – catch | catch > 0 with large YFT: standardised time series by year (black) with 97.5% confidence intervals (grey) compared to nominal (red)

3.4. Delta lognormal GLMM approach

The product of the three sub-models described above provided the standardised CPUE time series for free school sets by quarter (Figure 16) and by year (Figure 17). We considered the three components independent and calculate confidence intervals with:

$$Var(XY) = E(X^2Y^2) - (E(XY))^2 = Var(X)Var(Y) + Var(X)(E(Y))^2 + Var(Y)(E(X))^2$$

The formula was first applied to the product of the first (C1) and the second component (C2) then to the product of C1*C2 and the third component.

Purse Senners FSC CPUE - ind



Figure 16. Standardised CPUE for free school sets of yellowfin tuna category 2 & 3 (black line), with 97.5% CIs (grey,) and compared to nominal CPUE (red). Time series on a quarter basis.



Purse Seiners FSC CPUE - ind

Figure 17. Standardised CPUE for free school sets of yellowfin tuna category 2 & 3 (black line), with 97.5% CIs (grey,) and compared to nominal CPUE (red). Time series on an annual basis.

4. Discussion

In this paper we used the new approach developed in the Atlantic (Guéry *et al.* SCRS/2019/066_Rev) for CPUE standardization for the tropical tuna purse seine fisheries to account for the hierarchical structure of the tuna free

schools, for the non-randomised sampling and the numerous candidate variables linked to technological developments and evolving fishing strategies. A step forward compared to previous years was the inclusion of i) null sets, considered as presence of YFT FSC, ii) fishing days without set, considered as absence of FSC, iii) EU fishing agreement in the exclusive economic zones driving EU purse seine fleet presence in these areas, iv) time spent by centroid cell by boat by day to constrain detectability, v) the Gulland's index of fishing effort concentration to measure the extent to which a fleet has concentrated its fishing effort in areas with higher than average catch rates and, vi) piracy as a presence absence variable.

Despite the significant improvements in this analysis with respect to prior work, there are a number of different avenues for further improvement. Among the most interesting and pressing is the inclusion of environmental covariates. Environmental variability is undoubtedly a strong driver of tropical tuna fisheries, and, therefore, the inclusion of environmental covariates in our models is desirable. Nevertheless, this is challenging because it is often not clear if an environmental covariate primarily drives abundance (and therefore should vary when predicting standardized CPUE values) or is simply impacting catchability (and therefore should be fixed at a standard value when making predictions). For example, mixed layer depth undoubtedly impacts primary production, which ultimately impacts tuna stock size, but it also, influences the depth distribution of tunas and therefore their accessibility to surface purse-seine gear. One potential approach to resolving this issue is to split environmental variability into long-term climatological variability and short-term anomalies, the first being associated with overall abundance in the area and the second being associated with short-term changes in catchability.

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APPENDICES

DATA PER YEAR-QUARTER

Year_qu CPUE CPUE	
arter Stand CPUE (CI97.5) (CI97.5) SE CPUE SD	CV (%)
<u>1991_1</u> 2,081687909 1,419722686 2,743653131 0,295520189 9,624375231 4.	4,623351652
<u>1991_2</u> <u>1,617407446</u> <u>1,349332758</u> <u>1,885482133</u> <u>0,1196762</u> <u>3,515804089</u> <u>2</u>	2,173728147
<u>1991_3</u> 0,884662727 0,737763089 1,031562364 0,065580195 2,162833251 2	2,444811097
<u>1991_4</u> 1,187705031 1,031795114 1,343614948 0,069602642 2,238524764 1.	1,884748069
<u>1992_1</u> 2,519989454 2,157028981 2,882949927 0,162035925 6,14037289 2	2,436666107
<u>1992_2</u> <u>1,980658195</u> <u>1,658652244</u> <u>2,302664146</u> <u>0,143752657</u> <u>4,937995742</u> <u>2</u>	2,49310848
<u>1992_3</u> 1,083553671 0,762902791 1,404204551 0,143147714 4,57246726 4.	4,219880733
<u>1992_4</u> 1,439704922 1,225557786 1,653852058 0,0956014 3,429304316 2.	2,381949427
<u>1993_1</u> 2,50620358 2,144595206 2,867811954 0,16143231 6,339471741 2	2,529511885
<u>1993_2</u> <u>1,973023239</u> <u>1,715036612</u> <u>2,231009865</u> <u>0,115172601</u> <u>3,98429078</u> <u>2</u>	2,019383605
1993_3 1,079405937 0,915595417 1,243216456 0,073129696 2,712960221 2	2,513382713
1993_4 1,432097411 1,115314585 1,748880238 0,141420905 4,94147643 3	3,450516977
1994_1 1,153876512 0,832096504 1,475656519 0,143651789 5,126251331 4	4,442634267
1994_2 0,864921689 0,640360369 1,089483009 0,100250589 3,075024824 3	3,555263861
1994_3 0,472807174 0,413246485 0,532367864 0,026589594 0,926217088 1	1,958974268
1994_4 0,655545574 0,562300011 0,748791138 0,041627484 1,515587179 2	2,311947847
1995_1 1,818881595 1,582964343 2,054798847 0,105320202 3,965918079 2	2,180415751
1995_2 1,410878157 1,20201214 1,619744174 0,093243758 3,171711724 2	2,248040845
1995 3 0,77167783 0,639165841 0,90418982 0,059157138 2,269876614 2	2,941482214
1995 4 1,037560437 0,846205645 1,228915229 0,085426246 3,046972588 2	2,936669979
1996 1 1,875611016 1,617372631 2,133849402 0,115284994 4,374821122 2	2,332477834
1996 2 1,445171205 1,220123192 1,670219217 0,100467863 3,493154951 2	2,417121888
1996 3 0,79034799 0,664535717 0,916160263 0,056166193 1,770847159 2	2,240591717
1996 4 1,069080875 0,921107486 1,217054263 0,066059548 2,136425123 1	1,99837559
1997 1 1,207360357 1,01763862 1,397082093 0,084697204 3,210653533 2	2,659233853
1997 2 0,912446311 0,77496914 1,049923482 0,061373737 2,183846485 2	2,393397242
1997 3 0,498851185 0,432735517 0,564966853 0,029515923 0,943826715 1	1,892000548
1997 4 0,686605391 0,617319672 0,75589111 0,030931125 0,910838474 1	1,326582177
1998 1 1,085133362 0,929046966 1,241219758 0,069681427 1,659891168 1	1,529665594
1998 2 0,809933554 0,584240031 1,035627076 0,100756037 2,794086521 3	3,449772525
1998 3 0,442718277 0,373394978 0,512041575 0,030947901 0,859229284 1	1,940803734
1998 4 0,616173633 0,531603669 0,700743597 0,037754448 0,732173157 1	1,188257852
1999 1 2,110563877 1,775406028 2,445721727 0,14962404 5,257677735 2	2,491124667
1999 2 1,584209791 1,385735988 1,782683594 0,088604376 2,747254937 1	1,734148439
1999 3 0,86602296 0,765473438 0,966572482 0,044888179 1,021088789 1	1,179055101
1999 4 1,199261649 0,996952659 1,401570639 0,090316514 2,349551884 1	1,959165363
2000 1 2,223177582 1,743744438 2,702610727 0,214032654 6,763521238 3	3,042276646
2000 2 1,676523517 1,448029351 1,905017684 0,102006325 2,811827922 1	1,677177739
2000 3 0.916554823 0.743889555 1.089220091 0.077082709 1.871126628 2	2.041478132
2000 4 1.263957462 1.053835477 1.474079448 0.093804458 2.644895868 2	2.09255133
2001 1 2.372632776 1.939416748 2.805848804 0.193400013 6.266149392 2	2.641011056
2001 2 1,821966379 1,595329215 2.048603543 0.101177305 3.063089161 1	1,681199608
2001 3 0.996358434 0.850219086 1.142497782 0.06524078 1.614492321 1	1.62039309
2001 4 1.351840625 1.096361362 1.607319888 0.114053242 3.55535906 2	2.630013475
2002 1 1.911739522 1.603213847 2.220265197 0.137734676 4.382268938 2	2.292293949
2002 2 1.416817368 1.224435091 1.609199645 0.085884945 2.602033243 1	1.836533982
2002 3 0.774361218 0.705027882 0.843694553 0.030952382 0.679730329 0	0.877794901
$\begin{array}{ c c c c c c c c c c c c c c c c c c c$	2 415288512
2003 1 1.760254522 1.191350174 2.32915887 0.253975155 7.88298904 4	4.478323414
2003 2 1,311165091 1,065568212 1.55676197 0,109641464 3,02376717 2	2,306168149

2003 3	0,7166738	0,573557772	0,859789827	0,063891084	1,587001979	2,214399326
2003_4	0,999282689	0,681808036	1,316757342	0,141729756	4,250131973	4,253182827
2004_1	2,764170285	1,694769686	3,833570884	0,477410982	15,51236731	5,611943444
2004 2	2,105706012	1,757082254	2,45432977	0,155635606	4,586096001	2,177937459
2004 3	1,426336623	1,309635411	1,543037836	0,052098756	1,118246004	0,783998662
2004 4	2,018639969	1,786922095	2,250357844	0,10344548	3,228959033	1,599571535
2005 1	2,50056635	2,079211111	2,921921588	0,188105017	6,50800013	2,602610457
2005 2	1,776504656	1,609080856	1,943928456	0,074742768	2,292162921	1,290265642
2005 3	0,970320449	0,77185176	1,168789137	0,088602093	2,853942054	2,941236638
2005 4	1,411261482	0,853094409	1,969428555	0,249181729	8,234865827	5,835109887
2006 1	4,285675095	3,625524849	4,94582534	0,294709931	11,34296663	2,646716417
2006 2	3,146182391	2,865563909	3,426800874	0,125276108	3,955106505	1,257112911
2006_3	1,719291499	1,54741582	1,891167178	0,076730214	2,355999908	1,370331854
2006_4	2,42863085	2,20075842	2,65650328	0,101728763	2,940273968	1,210671424
2007_1	4,168768695	3,792232327	4,545305064	0,168096593	6,176113081	1,481519732
2007_2	3,094794409	2,842437536	3,347151282	0,112659318	3,517199477	1,136488895
2007_3	1,691504219	1,505093272	1,877915166	0,083219173	2,3795442	1,406762202
2007_4	2,365613118	2,118522967	2,61270327	0,110308103	3,75910418	1,589061267
2008_1	4,035484713	3,523178858	4,547790568	0,228707971	8,493589567	2,104725992
2008_2	2,931151758	2,670806696	3,19149682	0,116225474	3,254724906	1,110391128
2008_3	1,601521415	1,48022409	1,72281874	0,054150591	1,250744091	0,780972442
2008_4	2,283851547	2,032419925	2,53528317	0,11224626	2,953686666	1,293291882
2009_1	1,998283529	1,783451442	2,213115616	0,095907182	2,295807579	1,148889808
2009_2	1,370731417	1,2535702	1,487892634	0,052304115	1,040035297	0,758744772
2009_3	0,748302331	0,675858014	0,820746647	0,032341213	0,576894401	0,770937598
2009_4	1,122721197	0,996242988	1,249199407	0,056463487	1,081334329	0,963137003
2010_1	1,948998723	1,780524629	2,117472817	0,075211649	1,704520465	0,874562125
2010_2	1,362754955	1,19571805	1,529791859	0,074570047	1,858141463	1,363518405
2010_3	0,744155647	0,678485679	0,809825616	0,02931695	0,480767471	0,64605768
2010_4	1,097741373	0,929280719	1,266202027	0,075205649	1,70757852	1,555538091
2011_1	1,454370692	1,292672091	1,616069293	0,072186876	1,661678666	1,142541359
2011_2	1,032735033	0,917732527	1,147737539	0,051340404	1,34653827	1,303856485
2011_3	0,564071993	0,487625231	0,640518755	0,034128019	0,930719379	1,650001047
4	0,820761838	0,664341093	0,977182582	0,069830689	1,78507429	2,174899232
2012_1	2,259379117	1,952145829	2,566612404	0,137157718	3,983688765	1,763178537
2012_2	1,638979408	1,415459754	1,862499063	0,09978556	2,621777273	1,599640154
2012_3	0,895487235	0,808868331	0,982106139	0,038669153	0,961195305	1,073376892
2012_4	1,278474478	1,095381489	1,461567466	0,081737941	2,313273138	1,809401109
2013_1	1,332013456	1,111660555	1,552366357	0,098371831	1,803503196	1,3539677
2013_2	0,977643943	0,869082246	1,08620564	0,048465043	0,823217903	0,842042656
2013_3	0,534250455	0,472314608	0,596186303	0,027649932	0,434950837	0,814132834
2013_4	0,754813246	0,661218824	0,848407667	0,041783224	0,584735757	0,774676068
2014_1	1,921166864	1,686990947	2,155342781	0,10454282	3,222219369	1,677219938
2014_2	1,397770404	1,213951784	1,581589023	0,082061884	2,341421912	1,675111954
2014_3	0,76373281	0,689524949	0,83794067	0,033128509	0,903779785	1,183371689
2014_4	1,087495374	0,968311214	1,206679535	0,053207215	1,63295186	1,501571316
2015_1	1,438665659	1,188412949	1,68891837	0,11171996	3,161922106	2,197815792
2015_2	1,029700157	0,871026572	1,188373743	0,070836422	2,008916451	1,950972268
2015_3	0,562481262	0,499974834	0,624987691	0,027904656	0,792540787	1,409008336
2015_4	0,81270861	0,737166574	0,888250646	0,033724123	1,012115022	1,24536028
2016_1	1,523980724	1,359755547	1,688205901	0,073314811	2,199908767	1,443527948
2016_2	1,095642044	0,979022869	1,212261219	0,052062132	1,405443332	1,282757758
2016_3	0,598542899	0,51955404	0,677531757	0,035262883	0,947985846	1,583822726
2016_4	0,861384975	0,760662532	0,962107418	0,044965376	1,349230916	1,56635065
2017_1	0,950312295	0,785522728	1,115101862	0,073566771	2,242898121	2,360169529
2017_2	0,668788859	0,59545217	0,742125548	0,032739593	0,847799526	1,267663949
2017_3	0,365238609	0,332625938	0,397851279	0,014559228	0,340051368	0,931038944
2017_4	0,535691349	0,485464607	0,585918091	0,022422653	0,488108101	0,911174133

		Lower stand CPUE	Upper stand CPUE			
Year	Stand CPUE	(CI97.5)	(CI97.5)	SE CPUE	SD	CV (%)
1991	1,442865778	1,13981229	1,745919266	0,135291736	8,413860294	5,831353423
1992	1,75597656	1,428717901	2,08323522	0,146097616	9,935938496	5,658354855
1993	1,747682542	1,464206308	2,031158776	0,12655189	9,033371702	5,168771495
1994	0,786787737	0,615745331	0,957830144	0,076358217	5,0897071	6,468971055
1995	1,259749505	1,06301408	1,45648493	0,087828315	6,339917189	5,032680834
1996	1,295052771	1,102033055	1,488072488	0,086169516	5,897875086	4,554158113
1997	0,826315811	0,703771658	0,948859964	0,054707211	3,672794095	4,444782548
1998	0,738489706	0,595216036	0,881763376	0,06396146	3,141743462	4,254281997
1999	1,440014569	1,216104241	1,663924898	0,099959968	5,791192712	4,021620916
2000	1,520053346	1,234536449	1,805570244	0,127462901	7,044631895	4,634463595
2001	1,635699553	1,354155536	1,91724357	0,125689293	7,320639455	4,475540413
2002	1,296882743	1,092274665	1,501490821	0,091342892	5,102260812	3,934249908
2003	1,196844025	0,865847451	1,5278406	0,147766328	8,23357736	6,879407161
2004	2,078713222	1,601499171	2,555927274	0,213041987	11,74538154	5,650313575
2005	1,664663234	1,285869937	2,043456531	0,16910415	10,51688166	6,317723274
2006	2,894944959	2,532478222	3,257411695	0,161815507	10,27554358	3,549478048
2007	2,83017011	2,554994632	3,105345589	0,122846196	7,998802034	2,826261928
2008	2,713002358	2,402211722	3,023792995	0,13874582	7,951705083	2,93096136
2009	1,310009618	1,172152542	1,447866695	0,061543338	2,495530028	1,904970767
2010	1,288412674	1,14395397	1,432871379	0,064490493	2,84768314	2,21022596
2011	0,967984889	0,83524471	1,100725068	0,059259008	2,954068795	3,051771602
2012	1,518080059	1,320042652	1,716117467	0,088409557	4,797307432	3,160114911
2013	0,899680275	0,77509726	1,02426329	0,055617418	1,80512421	2,006406343
2014	1,292541363	1,142079075	1,443003651	0,067170664	3,940898775	3,048953703
2015	0,960888922	0,830868668	1,090909177	0,058044756	3,331581936	3,467187371
2016	1,01988766	0,902673696	1,137101625	0,052327663	2,975521975	2,91749973
2017	0,630007778	0,54306679	0,716948766	0,038812941	1,926611718	3,058076084

COEFFICIENTS OF THE COMPONENT 1 (POISSON DISTRIBUTION)

Generalized linear mixed model fit by maximum likelihood (Laplace Approximation) ['glmerMod'] Family: poisson (log)

Formula: num_sets_fsc ~ pays + age + num_sets_fob + cap_m3 + yr + quarter + Gulland + Piracy + (1 | c_bat) + (1 | eez_iso_3digit:fishing_access) + offset(searching_centroid) Data: D AIC BIC logLik deviance df.resid 491771.0 492152.1 -245847.5 491695.0 167301 Scaled residuals: Min 1Q Median 3Q Max -3.077 -0.701 -0.426 -0.109 158.277

Random effects:

Groups Name Variance Std.Dev.

c_bat (Intercept) 0.1700 0.4123

eez iso 3digit:fishing access (Intercept) 0.2308 0.4804

Number of obs: 167339, groups: c_bat, 76; eez_iso_3digit:fishing_access, 10

Fixed effects:

	Estimate	Std. Error	z value	$Pr(\geq z)$	
(Intercept)	-1.455056611	0.189162065	-7.69211635	1.447209e-14	***
pays4	0.289983807	0.124763035	2.32427663	2.011067e-02	*
age	-0.222187130	0.041097101	-5.40639420	6.430618e-08	***
num_sets_fob	-0.122979624	0.007238805	-16.98893943	9.916604e-65	***
cap_m3	0.003323342	0.065361909	0.05084524	9.594488e-01	
yr1992	0.265508787	0.026819172	9.89996221	4.164324e-23	***
yr1993	0.248218116	0.028249811	8.78654062	1.542359e-18	***
yr1994	-0.010560520	0.031435833	-0.33593892	7.369169e-01	
yr1995	0.039581475	0.034896595	1.13425034	2.566896e-01	
yr1996	0.111869781	0.038320180	2.91934382	3.507691e-03	**
yr1997	-0.134455518	0.043473929	-3.09278507	1.982877e-03	**
yr1998	0.079818017	0.048724378	1.63815363	1.013896e-01	
yr1999	0.350331581	0.049865223	7.02556931	2.131951e-12	***
yr2000	0.297152347	0.054004444	5.50236840	3.747229e-08	***
yr2001	0.365439682	0.058196634	6.27939550	3.398920e-10	***
yr2002	0.155093612	0.063558256	2.44018044	1.467993e-02	*
yr2003	0.065441822	0.067654619	0.96729275	3.333977e-01	
yr2004	0.307892365	0.071930349	4.28042363	1.865379e-05	***
yr2005	0.506415725	0.076167260	6.64873234	2.956280e-11	***
yr2006	0.683346438	0.081111237	8.42480607	3.613482e-17	***
yr2007	0.586736362	0.086416508	6.78963289	1.124192e-11	***
yr2008	0.567980500	0.091609778	6.19999866	5.646364e-10	***
yr2009	0.272891506	0.099019257	2.75594377	5.852306e-03	**
yr2010	0.333130986	0.105585126	3.15509390	1.604465e-03	**
yr2011	0.142649914	0.108246148	1.31782901	1.875609e-01	
yr2012	0.221640568	0.112288510	1.97384904	4.839891e-02	*
yr2013	-0.180220019	0.124498505	-1.44756774	1.477380e-01	
yr2014	0.055432572	0.122238626	0.45347837	6.502043e-01	
yr2015	0.070614050	0.126958091	0.55619968	5.780744e-01	
yr2016	0.207726363	0.132482507	1.56795314	1.168921e-01	
yr2017	0.074566447	0.138618818	0.53792442	5.906292e-01	
quarter2	0.099194593	0.010623318	9.33744027	9.869115e-21	***
quarter3	-0.666818086	0.013056726	-51.07085119	0.000000e+00	***
quarter4	-0.329046120	0.011017016	-29.86708318	5.268171e-196	***
Gulland	0.343772805	0.003702762	92.84226408	0.000000e+00	***
Piracy	-0.301251457	0.032041783	-9.40183199	5.362119e-21	***

Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

COEFFICIENTS OF THE COMPONENT 2 (BINOMIAL DISTRIBUTION)

Generalized linear mixed model fit by maximum likelihood (Laplace Approximation) ['glmerMod'] Family: binomial (logit)

Formula: yft_pos ~ pays + cap_m3 + annee_de_peche + trimestre + Gulland +
 (1 | numbat) + offset(nb_de_calees_pos)
 Data: D
Control: glmerControl(optimizer = "bobyqa", optCtrl = list(maxfun = 30000))

AIC BIC logLik deviance df.resid

138577.7 138903.9 -69254.9 138509.7 108523

Scaled residuals:

Min 1Q Median 3Q Max -9.5558 -0.9132 0.4826 0.8547 2.9116

Random effects:

Groups Name Variance Std.Dev. numbat (Intercept) 0.1207 0.3475 Number of obs: 108557, groups: numbat, 76

Fixed effects:

	Estimate	Std. Error	z value	$Pr(\geq z)$	
(Intercept)	-0.78859634	0.08063651	-9.7796436	1.376945e-22	***
pays4	-0.54628593	0.09979062	-5.4743216	4.391908e-08	***
cap_m3	0.10234305	0.05142459	1.9901577	4.657356e-02	*
annee_de_peche1992	0.19988941	0.04853933	4.1180913	3.820234e-05	***
annee_de_peche1993	0.22966532	0.04780664	4.8040460	1.554911e-06	***
annee_de_peche1994	-0.53981246	0.04792644	-11.2633552	1.990336e-29	***
annee_de_peche1995	-0.02726223	0.04635787	-0.5880821	5.564772e-01	
annee_de_peche1996	-0.13458560	0.04721599	-2.8504240	4.366099e-03	**
annee_de_peche1997	-0.42399658	0.04838378	-8.7631961	1.897911e-18	***
annee_de_peche1998	-0.59911435	0.05699020	-10.5125850	7.559265e-26	***
annee_de_peche1999	-0.52055772	0.05288527	-9.8431507	7.337622e-23	***
annee_de_peche2000	-0.45476172	0.05227425	-8.6995368	3.332416e-18	***
annee_de_peche2001	-0.18707932	0.05252919	-3.5614357	3.688324e-04	***
annee_de_peche2002	-0.69662233	0.05145806	-13.5376728	9.371097e-42	***
annee_de_peche2003	-0.62723936	0.05540503	-11.3209832	1.033067e-29	***
annee_de_peche2004	-0.30787895	0.05277719	-5.8335608	5.425682e-09	***
annee_de_peche2005	-0.72442679	0.04966080	-14.5874984	3.373423e-48	***
annee_de_peche2006	-0.27914295	0.04909777	-5.6854505	1.304683e-08	***
annee_de_peche2007	-0.12706957	0.04995947	-2.5434534	1.097627e-02	*
annee_de_peche2008	-0.42269004	0.05120268	-8.2552325	1.516068e-16	***
annee_de_peche2009	-1.24347970	0.05631289	-22.0816189	4.747731e-108	***
annee_de_peche2010	-0.95082038	0.05521889	-17.2191128	1.908918e-66	***
annee_de_peche2011	-0.73127552	0.05424321	-13.4814205	2.011954e-41	***
annee_de_peche2012	-0.44002651	0.05389472	-8.1645574	3.226158e-16	***
annee_de_peche2013	-0.28202258	0.08294983	-3.3999177	6.740613e-04	***
annee_de_peche2014	-0.40008467	0.05285174	-7.5699434	3.733868e-14	***
annee_de_peche2015	-0.62222737	0.05353587	-11.6226251	3.162667e-31	***
annee_de_peche2016	-0.56139277	0.05325504	-10.5415904	5.555139e-26	***
annee_de_peche2017	-0.85729625	0.05520764	-15.5285808	2.222442e-54	***
trimestre2	-0.29729973	0.01810809	-16.4180595	1.420371e-60	***
trimestre3	-0.30188522	0.01861215	-16.2197900	3.654421e-59	***
trimestre4	-0.03682949	0.01829317	-2.0132920	4.408392e-02	*
Gulland	0.83325468	0.01312961	63.4637790	0.000000e+00	***

Signif. codes: 0 '***' 0.001 '**' 0.01 '*' 0.05 '.' 0.1 ' ' 1

Linear mixed model fit by REML ['lmerMod']

Formula: log_capture ~ pays + cap_m3 + annee_de_peche + trimestre + Gulland +
 (1 | numbat) + offset(nb_de_calees_pos)
 Data: D
Control: lmerControl(optimizer = "bobyqa", optCtrl = list(maxfun = 40000))

REML criterion at convergence: 197377.5

Scaled residuals:

Min 1Q Median 3Q Max -4.8663 -0.6524 0.0623 0.7151 3.2019

Random effects:

Groups Name Variance Std.Dev. numbat (Intercept) 0.04239 0.2059 Residual 1.68247 1.2971 Number of obs: 58671, groups: numbat, 76

Fixed effects:

	Estimate	Std. Error	t value
(Intercept)	0.86935631	0.053627721	16.2109502
pays4	0.17126462	0.061091532	2.8034101
cap_m3	0.09443620	0.031488820	2.9990389
annee_de_peche1992	-0.12161704	0.036536279	-3.3286651
annee_de_peche1993	-0.11623846	0.036896560	-3.1503873
annee_de_peche1994	-0.41223674	0.038817010	-10.6200026
annee_de_peche1995	-0.16753299	0.036462574	-4.5946562
annee_de_peche1996	-0.18012338	0.038121714	-4.7249549
annee_de_peche1997	-0.28424807	0.041003065	-6.9323614
annee_de_peche1998	-0.54173273	0.049476805	-10.9492262
annee_de_peche1999	-0.17636691	0.045120895	-3.9087637
annee_de_peche2000	-0.09470548	0.041890059	-2.2608105
annee_de_peche2001	-0.18361027	0.041501909	-4.4241403
annee_de_peche2002	-0.01220757	0.044077921	-0.2769544
annee_de_peche2003	-0.03273176	0.041608818	-0.7866544
annee_de_peche2004	0.06341697	0.040859649	1.5520684
annee_de_peche2005	-0.28229520	0.039122698	-7.2156372
annee_de_peche2006	-0.14007822	0.039184459	-3.5748411
annee_de_peche2007	-0.13488259	0.040498468	-3.3305603
annee_de_peche2008	-0.01946548	0.041875591	-0.4648407
annee_de_peche2009	0.04424476	0.052081354	0.8495317
annee_de_peche2010	-0.22770881	0.048360239	-4.7085956
annee_de_peche2011	-0.45670980	0.045077741	-10.1316036
annee_de_peche2012	-0.24491700	0.043174700	-5.6726971
annee_de_peche2013	-0.44383635	0.061122142	-7.2614660
annee_de_peche2014	-0.25974942	0.042899524	-6.0548322
annee_de_peche2015	-0.45407207	0.043895173	-10.3444646
annee_de_peche2016	-0.56494518	0.043559871	-12.9693952
annee_de_peche2017	-0.74285066	0.046196950	-16.0800803
trimestre2	-0.26714294	0.015417330	-17.3274446
trimestre3	-0.10305813	0.015617528	-6.5988760
trimestre4	-0.22261198	0.014472152	-15.3820930
Gulland	0.58171446	0.009453955	61.5313353

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Quantifying the increase in fishing efficiency due to the use of drifting FADs equipped with echosounders in tropical tuna purse seine fisheries

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Numerous pelagic species are known to associate with floating objects (FOBs), including tropical tunas. Purse seiners use this behaviour to facilitate the capture of tropical tunas by deploying artificial drifting fish aggregating devices (dFADs). One major recent change has been the integration of echosounders in satellite-tracked GPS buoys attached to FOBs, allowing fishers to remotely estimate fishable biomass. Understanding the effects of this new technology on catch of the three main tuna species (yellowfin tuna, *Thunnus albacares*; bigeye tuna, *Thunnus obesus*; and skipjack tuna, *Katsuwonus pelamis*) is important to accurately correct for this change in catch-per-unit-effort (CPUE) indices used for stock assessments. We analysed catch data from the French purse seine fleet for the period 2010–2017 in the Indian Ocean to assess the impact of this fleet's switch to echosounder buoys around 2012. Results indicate that echosounders do not increase the probability a set will be succesful, but they have a positive effect on catch per set, with catches on average increasing by $\approx 2 - 2.5$ tonnes per set ($\approx 10\%$) when made on the vessel's own dFADs equipped with an echosounder buoy. Increases were due to a decrease in sets below ≈ 25 tonnes and an increase in those greater than ≈ 25 tonnes, with a non-linear transition around this threshold. This increase explains the considerable investment of purse seiners in echosounder buoys, but also raises concerns about bias in stock size estimates based on CPUE if we do not correct for this fishing efficiency increase.

Keywords: catch per unit effort, hyperstability, Indian Ocean, standardization

Introduction

Technological change has been a driving force behind increases in fishing efficiency worldwide for many decades now. For example, several studies have highlighted the substantial impact of new technology on catch rates of demersal fishes (Marchal *et al.*, 2007), tunas, and billfishes caught by pelagic longlines (Ward, 2008) and on the impact of global positioning systems (GPS) and plotter systems on the relative fishing power in a prawn fishery (Robins *et al.*, 1998). If the increase due to this new technology is not properly quantified and integrated into stock assessments, they will bias stock size estimates based on raw catch per unit

effort (CPUE) derived from captain's logbooks or observer data towards larger stock size, potentially leading to cryptic overfishing and hyperstability (Fonteneau *et al.*, 1999; Maunder and Punt, 2004). The process of correcting for this bias, known as CPUE standardization, is crucial to avoiding stock assessment errors that can lead to incorrect management decisions. However, for CPUE standardization to correctly account for bias, it is essential that data exist on when and where a given technology is being used and that the effect size be quantified via statistical analyses (Bishop, 2006). Unfortunately, this information is often unavailable. For example, in the case of the European tropical tuna purse

International Council for the Exploration of the Sea seine fishery that is the subject of this paper, it has been shown that new technologies have been introduced continuously over time, but the lack of a coordinated introduction of specific technologies across vessels and fleets and the absence of fine scale data (i.e. the date of the first use at the level of individual vessels and fishing trips) on these technologies has often made it difficult or impossible to fully account for their impact on catch rates (Gaertner and Pallarés, 2002). These issues represent a significant barrier to the use of purse seine CPUE data for stock assessments.

One aspect of tropical tuna purse seine fisheries that has seen significant technological change over recent decades is fishing on fish schools associated with floating objects (FOBs) (for a discussion of FOB terminology, see Gaertner et al., 2016; ICCAT, 2016). Tropical tuna purse seine fishing is dominated by two fishing modes: fishing on free schools (FSC) that are not associated with any object and fishing on FOBs. Though purse seiners have always fished on randomly encountered natural flotsam, since the early-1990s increasingly sophisticated technology has been used to deploy and track man-made FOBs, here referred to as drifting fish aggregating devices (dFADs) (Ariz et al., 1999; Hallier and Parajua, 1999). Today, the vast majority of FOB catch occurs on dFADs, and over 50% of the tropical tuna caught in purse seine fisheries are caught under FOBs (Fonteneau et al., 2013). This dominance is in no small part due to the use of tracking technology to remotely observe FOB movements in relation to oceanographic processes and historical catch patterns. Initially consisting of radio beacons attached to FOBs (both natural objects and dFADs), these have been replaced since the early-2000s with satellite-transmitting GPS buoys allowing fishers to determine an FOB location from virtually anywhere on the ocean (Torres-Irineo et al., 2014). The use of this tracking technology has broken the link between searching time and effective fishing effort for FOB sets as vessels can move directly towards a buoy, avoiding or significantly reducing searching time (Fonteneau et al., 1999). Most recently, echosounders have been integrated into GPS tracking buoys, remotely providing fishers information on the presence of tunas under the FOB, sometimes with a rough estimate of the fishable biomass (Baidai et al., 2020). This most recent innovation represents a major change in the way purse seiners fish that must be accounted for to properly evaluate trends in nominal CPUE values. Assessing these impacts is particularly important given that FOB fishing has been associated with increased juvenile catch, bycatch, ghost fishing, and potential disturbance to pelagic ecosystems (Hallier and Gaertner, 2008; Filmalter et al., 2013; Kaplan et al., 2014).

Here, we use a large dataset of fishing activity, catch rates, and GPS-buoy trajectories to examine the impact of echosounder buoys on catch rates for French tropical tuna purse seine vessels in the Indian Ocean from 2010 to 2017. French purse seiners switched to using almost exclusively echosounder equipped buoys after late-2011/early-2012, and in recent years 55-90% of fishing sets by this fleet are on FOBs. As this dataset spans the period before, during, and after this transition to echosounder buoys, we are able to estimate the impact of echosounders on catch rates. The key pieces of information essential to this estimation that were previously missing from catch-effort data are the ownership (i.e. whether the vessel has access to buoy information) and capabilities of the GPS buoy attached to the FOB associated with a fishing set. By combining data on FOB fishing sets with information on GPS-buoy models, trajectories, position reporting rates, and owner information, we are able to identify whether each set is on an FOB equipped with an echosounder buoy owned by the fishing vessel. Using this information, we can separate out catch on (i) FOBs not tracked by the fishing vessel from FOBs tracked by the fishing vessel (ii) with and (iii) without an echosounder, allowing us to assess differences in catch between the three.

Material and methods Data

Logbook catch and effort data and FOB trajectory data for French purse seine vessels operating in the Indian Ocean from 2010 to 2017 were used in this study. Data were provided by the fishing fleet to the IRD-Observatory of Exploited Tropical Pelagic Ecosystems (Ob7) based in the MARBEC research laboratory via an agreement with ORTHONGEL, the French frozen tuna producers' organization. Logbook data included the name and unique identifier of the fishing vessel, the date and geographical coordinates of each fishing set, the type of fish school (FOB or FSC), and the catch in tonnes for each of the three main tropical tuna species (skipjack, yellowfin, and bigeye). Null sets, i.e. sets for which the fishing vessel was unable to capture the associated fish school resulting in catch <1 tonne, were included in the dataset. Catch species composition data reported in logbooks were corrected based on port sampling using the T3 software (Pianet et al., 2000). Fishing sets were assigned to one of the four monsoon-based seasons (Escalle et al., 2017): 1 = December–March, 2 = April–May, 3 = June–September, and 4 = October–November.

FOB trajectory data consisted of the make and model of the tracking buoy, the unique numerical identifier of the buoy, and the date, time, and geographical coordinates of each buoy position. Tracking buoys typically emitted 1-2 positions per day, though the transmission rate was often increased to as much as 1 position every 15 min when vessels were actively tracking an individual FOB with the intention of carrying out a fishing set. Tracking buoys were assumed to be attached to an FOB, with the vast majority of these being dFADs in recent years (>90% based on observer data), though the data themselves do not indicate what the buoy is attached to. Buoys from the manufacturer Marine Instruments, which dominate French deployments from 2012 onward, also included echosounder data, if the buoy had one, and a list of vessel or company names with access to the buoy trajectory and echosounder data. Buoy positions were classified into onboard and in the water positions using a random forest algorithm based primarily on buoy speed, water temperature, and time step between successive positions (Maufroy et al., 2015; Imzilen et al., 2020).

French logbook and FOB trajectory data were considered to be exhaustive over the period of this study.

Determination of buoy ownership

For this study, we needed to determine for each FOB fishing set if the fishing vessel had access to the buoy trajectory and, if available, echosounder data. For fishing sets on FOBs equipped with more recent Marine Instruments buoys (in wide use by the French fleet after ≈ 2012), the primary method of making this determination was based on the vessel names found in both the logbook and the buoy trajectory data. Specifically, for all buoys located <11 km from the fishing set location within ±18 h of 00:00 UTC the day of the fishing set, we considered that the fishing vessel had access to the buoy position information if (i) the name of the vessel carrying out the set matched one of the vessel names associated with the trajectory data, (ii) at least 2 days of buoy trajectory data were available during the 5 days prior to the fishing set, (iii) the buoy trajectory for the 5 days preceding the fishing set was classified as being in the water \geq 90% of the time, and (iv) the minimum time between successive buoy positions was <1 h at some point over the 5 days preceding the set. The temporal window of ±18 h was needed as logbook data do not include the time of day of the fishing set. Specifically, ± 18 h was used to assure that at least one buoy position was included in the time period and based on the fact that the majority of fishing sets on FOBs occurs in the morning between 1 h before local sunrise and 10 h after local sunrise (based on anecdotal evidence and post hoc comparisons between logbook set locations and vessel trajectory data), which translates to no later than 18:00 UTC for the Indian Ocean. The spatial filter of 11 km was based on typical maximum distances buoys travel over a single day (Snouck-Hurgronje et al., 2018). The condition of 2 days of trajectory data was imposed to eliminate buoys deployed shortly before or after the fishing set was carried out. The final two conditions were imposed to assure that the buoy was actually in the water preceding the fishing set and not simply on board and that the vessel was actively trying to track the buoy by remotely lowering the time between successive transmissions of its position, respectively.

As Marine Instruments buoys were only in wide use by the French fleet after ≈2012, and from 2010 to 2013 Marine Instruments buoy trajectory data were often associated with company names instead of the names of individual fishing vessels, a second vessel-buoy association method based on separation distance was also used. If there was an FOB buoy trajectory that met all of the aforementioned conditions except the condition on matching vessel names, but that buoy was within 4 km of the set location, then it was also considered that the fishing vessel had access to the buoy data. The distance of 4 km was based on the observation that minimum separation distance for owned FOB sets meeting the prior association conditions based on vessel names was <4 km in the vast majority of cases (73%), whereas minimum separations for other types of sets (e.g. free school sets, foreign dFAD sets), for which there should be no valid buoy trajectory association, were generally much greater than this distance.

Based on these determinations and whether the associated buoy model possesses an echosounder, each FOB fishing set was placed in one of three set categories:

- *Foreign* (F), meaning that the fishing vessel had no access to buoy tracking information for the FOB upon which they fished.
- Owned echosounder (O E), meaning the fishing vessel had access to the buoy tracking data, but the buoy model was not echosounder equipped.
- *Owned* + *echosounder* (*O* + *E*), meaning the fishing vessel had access to tracking and echosounder data.

Note that the term *foreign* is used in the sense of not pertaining to the fishing vessel as opposed to not being from a specific flag (e.g. France) and the term *owned* may refer to buoys shared by multiple vessels, one of which is the vessel carrying out the set.

To assess the accuracy of our FOB ownership assignment methodology, we carried out the following tests: (i) the fraction of FSC sets matching our FOB ownership criteria was used to estimate our type I error as FSC sets should never be on an FOB; (ii) the fraction of FOB sets meeting the conditions to be on an owned FOB based on matching vessel names, but failing to meet the separation distance condition was used to estimate the type II error for this separation distance algorithm presuming that all matches meeting the vessel name condition are good matches; and (iii) the fraction of FOB sets classified as being on foreign FOBs but meeting all conditions for being on an owned FOB except for not having a time step of ≤ 1 h was used to evaluate the validity of using this as a condition for FOB ownership.

Data treatments

Preliminary analyses indicated that statistical analyses would benefit from the removal of some outlier data. First, it was found that one $5^{\circ} \times 5^{\circ}$ grid cell centred on 62.5° E and 17.5° N contained only two anomalously large fishing sets in a single year by a single vessel over the entire study time period. These two fishing sets were removed from the dataset.

Second, whereas most fishing vessels were active for the majority of the study time period, two vessels only fished in the Indian Ocean during 2010–2012 with low CPUE (Supplementary Table S1). Furthermore, very few of these sets were on FOBs with their own echosounder-equipped buoys, so data from these vessels were of little value for quantifying the effect of using echosounder buoys. Together the two boats represent 5.7% of the sets made between 2010 and 2017. It was therefore decided to remove these two vessels from the data.

Statistical analyses

Generalized additive models (GAMs), generalized linear models (GLMs), and beta regression models were developed to estimate the effect of having access to echosounder data on total catch and catch composition per successful set while controlling for potential sources of bias (see Table 1 for a full list of models). Catch per set was assumed to be Gamma distributed as this is a two-parameter positive definite distribution that is approximately normal for sufficiently small coefficient of variability. The proportion by weight of skipjack in a set was assumed to be beta distributed, and binomial distributions were used to model the relative probability of a set being on a foreign FOB vs. an owned FOB with echosounder as a function of set size category. Standard diagnostic plots of model residuals were used to assess the validity of these assumptions (Supplementary Figures S6–S11).

Six potential predictor variables were included in models:

- *Category*: Fishing set category (F, O E, O + E).
- Vessel: A categorical variable for the vessel unique identifier to account for random variability among vessels in terms of fishing efficiency.
- *Year*: A continuous variable from 0 to 7 corresponding to years 2010–2017.
- Season: A categorical variable from 1 to 4 for four Monsoonbased climatic seasons.
- Spatial parameters: 5° × 5° cell for GLM and beta regression models (*cwp55*); longitude (*lon*) and latitude (*lat*) for GAMs.
- *Size category*: For modelling fishing set category as a function of set size (model N1), a categorical variable for set size with 5-tonne bins, and a plus class for sets above 50 tonnes was used.

First GAM A1 (Table 1) was developed containing smooth terms for year by set category and for the interaction between longitude

Table 1. Models used to estimate the impact of echosounder buoys on catch of major tuna species.^{a,b}

ID	Depend. var.	Data distribution	Туре	Time period	Model equation
A1	Total catch	Gamma(link = identity)	GAM	2010-2017	te(lon, lat, by = season) + vessel + s(year, by = category)
A2	Total catch	Gamma(link = identity)	GAM	2012-2017	te(lon, lat, by = season) + vessel + s(year) + category
L1	Total catch	Gamma(link = identity)	GLM	2012-2017	cwp55 + vessel + season + year + category
A3	Total catch	Gamma(link = identity)	GAM	2012-2017	te(lon, lat, by = category: season) + vessel + s(year)
B1 ^c	Proportion of SKJ	Beta	Beta regr.	2010-2017	$cwp55 + vessel + season + category: year + year^{2} + year^{3} + year^{4}$
N1 ^d	Category F or $O + E$	Binomial	GAM	2012-2017	te(lon, lat, by = season) + vessel + s(year) + size class

^aModel equations are written using standard R syntax for the mgcv and betareg packages. The predictor variables are as follows: category = buoy category, vessel = unique vessel identifier, year = year of fishing set relative to 2010, season = monsoon seasons 1–4, lon, lat = longitude and latitude of set location, cwp55 = spatial parameter corresponding to 5° cell, and size class = total catch binned into 5 tonnes categories with a plus class for all sets >50 tonnes. ^bAfter testing different values for the *k* parameter in GAM smooths, these were fixed at 6 for s(year) and 13 for te(lon, lat).

^cThough for notational simplicity we have written *yearⁿ*, in reality orthogonal polynomials were used in model B1, limited to fourth order as higher-order polynomials were found not to be significant.

^dData for model N1 were limited to just set categories F and O + E.

and latitude by season (all other predictors were not smoothed). This model was used to identify spatial and temporal nonlinearity in catch per set, as well as identify temporal stability in the effect of set category on catch.

Based on the results of the first GAM A1, GAM and GLMs were then developed for the latter years (2012–2017) of the dataset for which the effect of set category on catch was relatively constant over time (models A2 and L1, respectively). Given this constant effect, a single smooth for year was used in model A2 for all fishing set categories. GLM L1 was included to confirm the robustness of results to model the formalism and description of spatial variability.

GAM A3 was developed to look at spatial inhomogeneity in the effect of fishing set category (and particularly O + E data) on catch per set. In this model, the interaction between set category and space was studied by estimating a different smooth for longitude \times latitude for each fishing set category and season combination.

Assuming that skipjack is the main target species for FAD fishing, changes in the proportion of skipjack by set for the period 2010–2017 were examined using beta regression (model B1). In this model, an interaction between set category and year was included to assess temporal changes in species composition as a function of set category. Orthogonal polynomials of the year predictor variable of up to fourth order were used to account for non-linear temporal variability in skipjack proportion.

Finally, to assess changes in the relative frequency of sets with different catch sizes when echosounders are in use, a binomial GAM was developed with a categorical catch size variable with 5-tonne bins and a plus size category for sets over 50 tonnes as a predictor and set category as the predicted variable (model N1). This model was limited to set categories *F* and O + E.

Statistical analyses were carried out using R version 3.6.3 (2020-02-29) (R Core Team, 2020). GAMs were estimated and visualized using the mgcv package version 1.8.33 (Wood, 2011, 2019) and the mgcViz package version 0.1.6 (Fasiolo *et al.*, 2020), respectively. Beta regression was done with the betareg package version 3.1.3 (Cribari-Neto and Zeileis, 2010).

Results

Accuracy of buoy ownership assignments

The type I error of our FOB ownership assignment methodology as assessed by the fraction of FSC sets, which should never truly be on an owned FOB, that nevertheless meet the conditions for being considered to be on an owned FOB was 4.8% out of 6065 FSC sets examined.

The type II error of the set-trajectory separation distance condition for being considered to be on an owned FOB was assessed by estimating the fraction of FOB sets meeting the ownership conditions based on matching vessel names (and, therefore, considered to be a good assignment) that fail to meet the separation distance condition. 15.8% of the 2610 FOB sets classified as on owned FOBs based on matching vessel names fall into this category. Though larger than the overall type I error, it is important to note that only 1.1% of the 7418 FOB sets classified as foreign do meet the base conditions of having a nearby (i.e. $\leq 11 \text{ km}$) buoy trajectory with at least 2 days of trajectory data over the 5 days preceding the set, 90% of which is classified in the water, and a minimum time step of $\leq 1 \text{ h}$. Therefore, this error only applies to a small percentage of the data.

Finally, only 0.5% of the potential FOB set-buoy trajectory matches that have matching vessel names and meet the base conditions for being a good match (e.g. 2 days of in water trajectory data) have a minimum time step of >1 h. This indicates that French purse seiners essentially always reduce the emission frequency of buoys when they approach an FOB, supporting the imposition of this condition for considering an FOB set to be on an owned FOB.

Preliminary analyses

There were no consistent differences between the proportions of null FOB sets for the three set categories (Figure 1; no significant differences between set categories found by a binomial GLM). The percentage of null sets generally varied between 6 and 10% for all three set categories except for a couple of year-category combinations for which very little data was available (e.g. 2014 O - E; Figure 1).

In each of the three set categories, the number of FOB sets has varied considerably over years (Figure 2). Though the proportion of (positive) FOB sets for which vessel's had access to buoy tracking information was $\approx 25 - 30\%$ in the early part of the time series, this proportion has increased dramatically since 2015, reaching $\approx 50\%$ in 2017. This change is entirely due to a large increase in the number of sets on the vessel's own echosounder-equipped buoys (solid curve in Figure 2a). The use of buoys without echosounders decreased steadily over the time series,



Figure 1. Proportion of null FOB sets from 2010 to 2017 for each set category. The set categories are: F = foreign FOB; O - E = owned buoy without echosounder; O + E = owned buoy with echosounder.

disappearing entirely after 2014 (dotted curve in Figure 2a). It must be noted that the number of FOB sets per vessel per year did not show strong or consistent pattern despite variations over the study period (Figure 3). The number of FOB sets appears to be marginally higher for 2010, coincident with the impacts of Somali piracy on the fishery, and for 2016–2017, coincident with both a gradual shift away from fishing on free schools after 2014 (not shown) and the imposition of a quota on yellowfin tuna in 2017 that led to increased fishing on FOBs. The increase in the total number of FOB sets in the Indian Ocean (Figure 2a) was, therefore, primarily due to the entry into the fishery of five vessels in 2014 (Supplementary Tables S1–S3).

Mean catch per set varied over years between 22 and 31 tonnes for all year-set category combinations for which reasonable amounts of data are available (Figure 4; median values in Supplementary Figure S1). From 2012 onward, mean catch for O + E sets is consistently higher than that for *F* sets, averaging 2.6 tonnes more catch per set. This corresponds to a 10.6% increase in catch per set for O + E sets over *F* sets.

Though catch species composition was globally similar between the different set categories, there does appear to be an interaction between set category, species composition, and year (Figure 5). Over the study time period, the fraction of skipjack in O + E sets increased relative to that of F sets, being initially lower than that for F sets, but eventually surpassing that of F sets (solid vs. dashed curves in Figure 5a). Conversely, the fraction of yellowfin decreased over time for O + E sets relative to that of F sets (solid vs. dashed curves in Figure 5b).

With the objective of identifying any potential interaction between set category and the area explored by fishers, we analysed the spatial distribution of sets for each set category (Supplementary Figure S4). Though there is interannual variability in the spatial distribution of fishing sets, there is no clear evidence of consistent differences in the spatial distributions of sets between set categories. Apart from the disappearance of O - Esets after 2015, results for all time period-set category combinations have approximately the same overall spatial extent and



Figure 2. Absolute number of positive FOB sets for each set category (a) and the proportion of all positive FOB sets that are in each category (b) for the period 2010–2017. The set categories are: F = foreign FOB; O - E = owned buoy without echosounder; O + E = owned buoy with echosounder. In (b), the proportions are cumulative with O + E representing the proportion of sets on owned buoys possessing echosounders and O indicating the total proportion of sets on owned buoys with *and* without echosounders. The proportion that *F* sets represent is one minus that for O sets. Note that there were no O + E sets in 2010, and from 2015 onward there were no O - E sets.

similar relative densities of sets per cell. This is also true for median catch per set (Supplementary Figure S5).

Statistical analysis

To eliminate the possibility that observed differences in catch per set with and without access to echosounder data are due to spatial, seasonal, or fishing vessel differences between *F* and O + E sets, we ran a series of GAM and GLM statistical models including these variables in addition to set category (see Table 1 for details). Results from the initial GAM A1 based on the entire 2010–2017 time series indicate that all of these factors have a significant impact on catch per set (Table 2). In particular, though there is interannual variability in catch per set as a function of the set category, catch per O + E set is relatively stable after ≈ 2012 (Figure 6). We, therefore, used the time period 2012–2017 for all



Figure 3. Boxplots of the number of FOB sets per vessel per year. Boxplots only include vessels for which at least one set was carried out in the given year. Whiskers on the boxplot represent the extremes for the given year.



Figure 4. Mean catch per set between 2010 and 2017 for each of the three set categories. The set categories are: F = foreign FOB; O - E = owned buoy without echosounder; and O + E = owned buoy with echosounder.

other models except model B1 used to assess change in skipjack proportion in the catch per set.

Results for GAM A2 and GLM L1 for the period 2012–2017 were consistent with exploratory analyses, indicating a significant increase in catch per successful set of \sim 2.0 to 2.5 tonnes per O + E set relative to F sets (see the coefficient of the O + E terms in Table 3). Spatial, temporal, and vessel terms did explain significant variance in catch per set (Supplementary Tables S4 and S6).

The interaction between the set category and spatial zone (model A3) was seasonally variable and relatively weak over much of the core purse seine fishing grounds (Figure 7). Nevertheless, catch per O + E set tended to be larger than that for

F sets in the core FOB fishing areas east of Somalia for seasons 1 and 4 and east of Tanzania for season 2, and the reverse is true for the zone east of Tanzania for season 1. Patterns are more complex and variable for other seasons and zones (specifically areas less explored by the French purse seiners).

Beta regression model B1 confirmed that even after accounting for spatial and vessel differences, there was a significant interaction between year and set category on the proportion of catch that is skipjack (Table 4), with the skipjack proportion increasing over time for O + E sets relative to F sets.

The use of FOBs equipped with echosounder buoys has also significantly changed the set size distribution. Binomial GAM N1 indicates that O + E sets since 2012 have a small proportion of sets below the mean set size of ≈ 25 tonnes relative to *F* sets and a large proportion of sets above this size (Figure 8). The transition from negative to positive differences is abrupt, suggestive of a threshold effect for the impact of having access to echosounder data on catch per set.

Discussion

Our results convincingly demonstrate how and when the use of echosounder buoys information has increased catch per FOB set in the French Indian Ocean tropical tuna purse seine fishery. Sets on FOBs for which fishers have access to echosounder data catch on average $\approx 2.0 - 2.5$ tonnes more tuna than sets on FOBs for which fishers do not have this information, either because the tracking buoy does not belong to the vessel or because the tracking buoy does not have an integrated echosounder. This increase corresponds to a $\sim 10\%$ increase in catch per set. Catch per set on FOBs with echosounder buoys is fairly consistent after 2012. Before 2012, there is evidence of a period during which French purse seiner skippers appear to have been learning how to take advantage of buoy echosounder data, though this period is also coincident with improvements in buoy echosounder technology that may have played a role. Though a 10% increase in catch may seem relatively moderate, 2.0-2.5 tonnes of tuna has an approximate cannery sale value of US\$2200-2800 (https://www.undercur rentnews.com/data/prices/#/skipjack_seychelles&start=0&end=5; the value quoted is an approximate mean for the period September 2019–February 2020). Given that a buoy is used on average three or four times before being lost or retired, the gains far outweigh the cost of an echosounder buoy (\approx US\$1000–1500; Lopez et al. 2014) by approximately US\$5000-7000. Furthermore, echosounder buoys also likely represent a gain in search time as they permit remotely determining the presence or absence of a tuna school under an FOB (Baidai et al., 2020). This gain may increase over time due to advances in research. For example, based on differences in the acoustic response between skipjack and bigeye, Moreno et al. (2019) highlighted the interest of predicting the species proportion in the tuna school for tuna conservation.

The use of echosounder-equipped tracking buoys is coincident with an increase in the proportion of FOB sets on buoys to which the vessel has access to tracking and echosounder information. The proportion of sets directed at a vessel's own FOBs and the increasing trend in this proportion are consistent, though on the high end, of prior, more approximate estimates of FOB fishing strategy (Snouck-Hurgronje *et al.*, 2018). This increasing trend highly suggests that the benefits of using this technology have been sufficient to cause fishers to change fishing strategy from one in which random encounters of foreign FOBs and free



Figure 5. Average proportion of catch in each of the three major species of tunas as a function of set category and year. The tuna species are (a) SKJ = skipjack, (b) YFT = yellowfin, and (c) BET = bigeye. The set categories are: F = foreign FOB; O - E = owned buoy without echosounder; O + E = owned buoy with echosounder.

Table 2. ANOVA table for a series of nested GAMs culminating in model A1. $\!\!\!^{\rm a}$

	Resid.	Resid.			
	Df	Dev	df	Deviance	Pr(>Chi)
NULL	10 265	8 282	-	-	-
te(lon, lat, by $=$ season)	10 008	7 585	257.47	697.4	0.0000***
vessel	9 992	7 553	15.83	31.5	0.0023**
s(year, by $=$ category)	9 973	7 482	19.18	71.2	0.0000***

^aSee Table 1 for model details.

**0.001-0.01

***0-0.001

schools were important components of overall fishing effort to a strategy increasingly directed towards the vessel's own FOBs for which the fisher has access to echosounder information. The potential contribution of this change to the recent increase in dFAD deployments (Lopez *et al.*, 2017; Imzilen *et al.*, 2020) is an important avenue for future research.

Increases in catch per set when using echosounder buoys occur in a non-linear fashion consistent with a threshold detection effect for fish schools. When using echosounder buoys, skippers tend not to set on small schools less than $\approx 20 - 25$ tonnes which results in an increase in the number of sets catching more than this amount, with a rapid, non-linear transition between these two states (Figure 8). Prior analyses have shown that echosounder buoys are effective at detecting the presence of fish schools above this size but provide only relatively weak information on the absolute size of fish schools (Baidai *et al.*, 2020). Our results are consistent with these observations, showing relatively homogeneous changes in the relative frequency of sets above and below this set size, with a narrow transition between the two school size categories.

We do not have a clear explanation for the observed increase in the relative proportion of skipjack in O + E FOB catches over time. One hypothesis is that as fishers became more accustomed to the new technology, they gradually became more efficient at identifying large schools of skipjack tuna, the primary target for FOB fishing. Recent concern regarding the status of yellowfin stocks in the Indian Ocean may also play a role, though the timing is off with concern increasing after about 2016 and a yellowfin quota being imposed in 2017. This trend merits further examination in future research, e.g. by comparison with results for the Atlantic Ocean and other purse seine fleets.

Though we did not observe strong spatial inhomogeneities in the core zone of purse seine fishing with respect to the impact of echosounder buoys on catch per set, there are hints of regional trends in the effectiveness of this technology for detecting fish schools. Catch on echosounder-equipped buoys belonging to the fishing vessel appears to be higher in vicinity of the Somali upwelling zone and lower in the northern Mozambique Channel and off Tanzania. These two areas are quite different in terms of productivity, fishing season, and fishing strategy. Monsoondriven upwelling off Somalia, roughly occurring from July to October, sustains essential nursery habitat for all species of tropical tuna (Kaplan et al., 2014). During this season, fishers concentrate the vast majority of their effort in this area and echosounder buoys appear to be highly effective at increasing FOB catch rates. By contrast, the Mozambique Channel is an important area for both FOB and free school fishing somewhat earlier in the year (March-June) (Kaplan et al., 2014). This area is characterized by eddy circulation and increased turbid due to river outflows



Figure 6. Effect of the interaction between year and set category estimated from GAM A1 (see Table 1). The panels are from top to bottom: (a) F = foreign FOB; (b) O - E = owned buoy without echosounder; and (c) O + E = owned buoy with echosounder. Vertical red lines indicate the range of years for which we have data for the set category corresponding to a given panel.

Table 3. Regression coefficients related to set category for GAM A2 and GLM ${\rm L1.}^{\rm a}$

	GAM A2		GLM L1	
	Estimate	Pr(> t)	Estimate	Pr(> <i>t</i>)
Intercept (cat. F)	22.00	0.0000***	21.7	0.0000***
Category $O - E$	-2.26	0.2241	0.2	0.9253
Category $O + E$	1.98	0.0001***	2.5	0.0000***

^aSee Table 1 for model details.

***0-0.001.

(Schott *et al.*, 2009). The interaction between anticyclonic and cyclonic eddies generates high dynamical and complex barriers consisting of multiple fronts at different scales favourable to the production and aggregation of organic matter (Tew-Kai and Marsac, 2009). This region is also characterized by a relatively high rate of occurrence of natural FOBs (Maufroy *et al.*, 2015). Though it is unclear exactly how these regional differences might impact the effectiveness of echosounder buoys, we speculate that water turbidity and stability of environmental conditions may impact both the quality of the echosounder signal and fishers capacity to accurately interpret echosounder data.

Our results have significant consequences for stock assessment. FOB catches provide important data for estimating the trend in the abundance of skipjack tuna and juvenile population sizes for yellowfin and bigeye tuna. The recent use of echosounder buoys is likely to cause positive bias in nominal CPUE series, potentially contributing to hyperstability and failure to detect overexploitation if this bias is not properly accounted for. For example, using predictions from model A2 where we have artificially

Season 1 Season 2 10 10 Latitude Latitude 0 0 -10 -10 Diff (t) -20 -20 70 80 60 60 70 80 20 50 50 40 40 Longitude Longitude 10 Season 3 Season 4 0 -1010 10 -20 -30 Latitude Latitude 0 0 10 -20 .20 70 40 60 70 80 40 50 60 80 50 Longitude Longitude

Figure 7. Difference between the smoothed spatial effects for fishing set categories *F* and O + E for each season from GAM A3 (see Table 1). Colours indicate the difference between the two effects in tonnes, with positive values indicating higher catch per set for O + E sets. Differences have been constrained to the range -30 to 30 tonnes so as to avoid anomalously large predictions in areas with little data (i.e. dark blue and bright yellow regions in peripheral areas for some reasons). Black solid and dotted contour lines indicate differences of +5 and -5 tonnes, respectively. Red solid and dotted contour lines indicate significant differences at the 0.05 and 0.01 levels, respectively, as determined by the plotDiff function of the mgcViz package.

Table 4. Regression coefficients related to set category and year for beta model B1.^{a,b}

	Estimate	Pr(>z)
Intercept (cat. F)	0.05	0.5701
Year	7.14	0.0000***
Year ²	6.59	0.0000***
Year ³	-7.76	0.0000***
Year ⁴	-1.52	0.0935
Category $O - E$	0.08	0.4796
Category $O + E$	0.00	0.8112
Year: category $O - E$	2.82	0.7120
Year: category $O + E$	15.39	0.0000***

^aSee Table 1 for model details.

^bThough for simplicity we use the notation of raw polynomial terms, the Yearⁿ terms actually represent the *n*th order orthogonal polynomials. ***0-0.001

standardized all sets to be category F, we estimate that the use of echosounders increased total FOB catch by the French fleet between 1.7 and 4.0% per year over the period 2012–2017. Though relatively small, this amount could be critical for the management of fully exploited tuna species. Furthermore, the impact of echosounders is likely to grow over time as purse seiners increasingly target their own FOBs and echosounder technology improves, highlighting the importance of accounting for these changes.

The question is how to apply this knowledge to the standardization of CPUE data for the ensemble of tropical tuna purse seine fleets. Ideally, one would collect data on the provenance and characteristics of FOBs associated with fishing sets and directly integrate this information into CPUE standardization models. Though fleets are beginning to collect more data on FOB use, these data will likely never be available for historical data and may not directly address questions of buoy ownership. Back calculating ownership using FOB tracking buoy trajectories as we have done in this paper is a possibility, but, for example the Spanish purse seine fleet began using echosounder buoys in the 2000s (Lopez et al., 2014) long before trajectory data were regularly saved and made available to scientists. Nevertheless, if mean data per fleet, ocean, and year on the probability of fishing on an FOB equipped with echosounder buoys can be obtained (for example by extrapolating from known fleet purchases of echosounder buoys and recent estimates of the rate of fishing on owned FOBs), then applying a constant negative corrective factor to all FOB catches may be the best available solution given the relatively consistent impact of echosounder buoys on catch per set over the period of our study (\approx 2.0 tonnes per set).

Though we consider that our results are approximately applicable to all purse seine fishing fleets, and in particular the Spanish and Seychelles fleets, ideally this would be validated by carrying out similar analyses for each fleet. The Spanish fleet currently



Figure 8. Relative probability of sets being of type *F* vs. O + E as a function of set catch size category for data from 2012 to 2017. Results show the coefficients of the terms corresponding to the different catch size categories in model N1 (see Table 1). Positive values along the *y*-axis indicate that a set of a given size is more likely to be of category O + E than category *F*. As we are only concerned with relative probabilities, coefficients have been centred by removing the mean of all the coefficients. Note that the final size category is a plus class containing all sets >50 tonnes.

uses a wider variety of buoy models and manufacturers (Lopez *et al.*, 2016), and the extent to which technological differences between different buoy models impact the effectiveness of echosounder information for identifying large fishing schools is currently poorly understood. We, therefore, hope that these analyses can be carried out in the near future, though data access limitations and lack of historical data for the Spanish fleet may represent significant barriers. In the absence of these analyses, our results represent the best available data regarding the impact of this important technological change on purse seine catch per set and we strongly encourage the integration of our results in future stock assessments.

Supplementary material

Supplementary material is available at the *ICESJMS* online version of the manuscript.

Data availability statement

Given the confidential nature of the data used in this article, requests for data access should be addressed directly to the Ob7 (https://www.ob7.ird.fr/en/) pelagic ecosystem observatory using the following e-mail address: adm-dblp@ird.fr.

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