

Assessing the effectiveness of LED lights for the reduction of sea turtle bycatch in an artisanal gillnet fishery – a case study from the north coast of Kenya

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Abstract

Artisanal gillnet fisheries exist throughout the world's oceans and have been responsible for high bycatch rates of sea turtles. Three sites on the north coast of Kenya, i.e. Watamu, Ngomeni, and Bwana Said, were studied with the overall objective of assessing the effectiveness of LED lights in the reduction of sea turtle bycatch in the bottom-set gillnet fishery. A total of 10 boats with pairs of control and illuminated nets were deployed during the study, with 56 turtles caught in control nets, while 30 were caught in illuminated nets. The mean catch per unit effort (CPUE) of target species was similar for both control and illuminated nets. In contrast, the mean CPUE of sea turtles was reduced by 64.3% in illuminated nets. This statistically significant decrease ($p < 0.04$) in sea turtle catch rate suggests that net illumination could be an effective conservation tool. Some useful data on fish catch rates with and without LED lights were also obtained, and interviews with fishermen suggested that they believe that the lights are effective at reducing marine turtle bycatch in their gill nets when set at night. The issues associated with implementing the use of LED lights included increased net handling times, equipment costs, and limited awareness among fishermen regarding the effectiveness of this technology. These challenges need the support of other stakeholders, especially national government, so as to implement this strategy of reducing turtle bycatch more widely.

Keywords: mortality; sea turtles; bycatch; small-scale fisheries; gillnet; LED lights

Introduction

Five species of sea turtle have been documented from Kenyan waters (Frazier, 1975). These are the green turtle (*Chelonia mydas*), hawksbill turtle (*Eretmochelys imbricata*), loggerhead turtle (*Caretta caretta*), olive ridley turtle (*Lepidochelys olivacea*) and leatherback turtle (*Dermochelys coriacea*). Some of these sea turtles are known to nest on sandy shores in Kenya, while most forage in a diverse range of marine habitats, which include coral reefs, seagrass meadows, and mangrove swamps off Kenya (Wamukoya *et al.*, 1997; Okemwa *et al.*, 2004).

As with global sea turtle populations, those in Kenya also face increasing anthropogenic threats that put the populations at risk (IUCN, 2007). Typical threats include entanglement in fishing gear, poaching and illegal trade of eggs, meat, and shells, coastal development

which destroys nesting habitat and disorients sea turtles, plastic pollution and other marine debris, and global warming. The primary risk to sea turtles is entanglement or capture in the gear used in commercial and artisanal fisheries (IUCN, 1995; Spotila *et al.*, 2000; Lewison *et al.*, 2004; Alfaro-Shigueto *et al.*, 2010). In Kenya a significant number of sea turtles, among them endangered green sea turtles, are killed each year from the purse seine and gillnet fisheries (Wamukota, 2009; Watamu Turtle Watch, 2014; Fisheries Department, 2014). Although it is illegal to capture sea turtles in Kenya, the incidental capture of these animals remains a major threat to turtle populations in the region (WWF, 2009; KMFRI, 2016).

Bycatch in fisheries has been recognised as a global threat to sea turtles (WWF, 2009). The interaction of

small-scale coastal gillnet fisheries with sea turtles is documented to equal, or in some cases exceed, interactions with industrial pelagic fisheries (FAO, 2009). It has been estimated that over 70% of sea turtle interactions with fisheries worldwide end-up in capture (Molony, 2005; FAO, 2009; Alfaro-Shigueto *et al.*, 2011) and this poses a significant source of mortality for sea turtles (Hays-Brown, 2003; WWF, 2016). Watamu Turtle Watch in Kenya reported more than 273 turtle mortalities per annum related to interactions with fisheries within Watamu coastal fishing of Watamu (pers. comm, September 20, 2015). This rate of incidental capture and mortality poses a great threat to sea turtle management and population recovery and has thus been the focus of recent conservation work. Several studies have suggested that illuminating fishing nets with LED lights can reduce sea turtle capture by up to 40% without any significant impact on the catch of targeted fish (Wang *et al.*, 2010, 2013). These studies used either light-emitting diode (LED) light sticks or chemical light sticks to illuminate portions of nets.

To help limit the negative impacts of fisheries, bycatch reduction technologies (BRTs) have been developed for a limited number of fisheries (Cox *et al.*, 2007). Much effort has focused on the use of circle hooks in longline fisheries (Gilman *et al.*, 2006; Serafy *et al.*, 2012) and the use of Turtle Excluder Devices (TEDs) in shrimp trawl fisheries (Crowder *et al.*, 1994, 1995; Watson *et al.*, 2005; Lewison and Crowder, 2006; Read, 2007; Jenkins, 2011). However, the development of bycatch mitigation measures for gillnets, one of the most ubiquitous gear types, has been comparatively slow (Melvin *et al.*, 1999; Gilman *et al.*, 2006).

Kenya has an estimated 4,450 boats in its fisheries (KMFRI, 2016; Fisheries Department, 2016) that fish six days per week for target species such as the blue shark (*Prionace glauca*), flounder species (*Paralichthys* spp.), guitarfish (*Rhinobatos planiceps*), kingfish (*Scomberomorus cavalla*), and shortfin mako shark (*Isurus oxyrinchus*). Since five species of sea turtles make use of the same waters as the fishing vessels (Amiteye, 2002), it is critical and imperative to quantify sea turtle mortality and assess the potential of LED lights in reducing sea turtle captures as a bycatch reduction tool.

Clearly, protecting Kenya's sea turtles is extremely important for the global recovery of the species. This study sought to 1) quantify the present sea turtle capture rate in the gill net fisheries, 2) compare capture rates in gill net fisheries with unmodified and

modified (gear illumination with LED lights) fishing gear, and 3) compare catch quantity and composition of target fish species in unmodified and modified fishing gear. It was hypothesised that LED lights would reduce sea turtle bycatch, and would not have any impact on the target species fished.

Methods

The Kenyan coast spans approximately 600 km in a north-north east to south-south west direction, between 1° and 4° S on the Indian Ocean (UNEP, 1998; Okwema, *et al.*, 2004). The study was undertaken between December 2016 to December 2017 at Watamu, Ngomeni and Bwana Said landing sites, approximately 105 - 150 km north of Mombasa. The sites host a large gillnet fishery and an abundance of sea turtles.

Sensitization of fishermen about the project was undertaken at the start of the research period, and the Split-Block Sampling Design (Fisher, 1925; Box *et al.*, 2005) used in data collection. Nets were operated as pairs so as to ensure the experimental and control nets were carried by the same boats to minimize the errors. The catch from the nets were categorized into three groups: target species (fish sold), bycatch (discarded fish), and other (catch kept by the fishermen for consumption or retained for bait in other unrelated fisheries). The turtle species, curved carapace length (CCL; notch to tip (cm)), the rate of turtle bycatch, and fate of by-catch was recorded. Live sea turtles were released in accordance with internationally recognized guidelines (Epperly *et al.*, 2004). Turtle catch-per-unit-effort (CPUE) for each net was calculated as the number of turtles captured/([net length/100 m] x [net soak time/12 h]). Data were also collected on fishing method, gear type, design and operation, target fish catch composition, and weight. The total target species CPUE was calculated as the number of individuals of target species/([net length/400 m] x [net soak time/12 h]).

Out of 30 boat captains identified at the start of the study, 10 were randomly selected to participate in the net trials. Each of the boats was equipped with bottom-set gill nets. Six boats at Bwana Said, two boats at Watamu and two boats at Ngomeni Participated in the study, based on the relative importance of each landing site. After the project the boat captains were interviewed through a questionnaire, primarily to determine the effect of the LED lights.

The fishing boats were provided with paired sets of gillnets (1 and 2) at the start of the study. Group 1 nets

(control) were to fish without any modification, and Group 2 nets were fitted with gear modification/illumination (Deep Drop LED Fishing light 2,100 ft Green) (Fig. 2A) placed every 15 meters along the gillnet float line. The gillnets were made of multifilament twine (Fig. 2B) and were composed of multiple net panels 50

Trained observers/fishermen recorded all catch and by-catch information. Plastic flipper tags were fixed to all sea turtles captured to determine recapture rates. Support items were provided for fishing boats used in the research and observers were given free LEDs as motivation to sustain interest in the research.

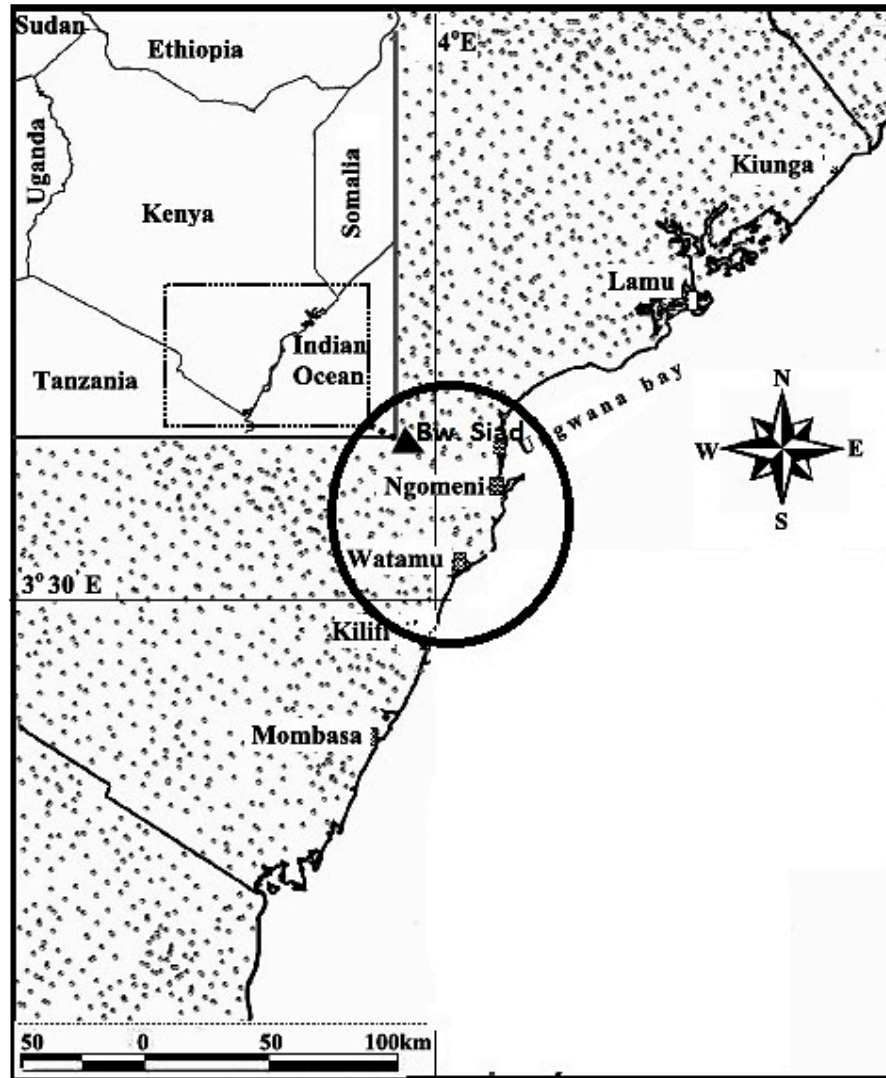


Figure 1. Map of the study area; Watamu, Ngomeni, and Bwana Said landing sites

m long by 3 m high. The number of gillnet panels set each evening varied, depending on the fishing crew, but averaged 5 panels per night. Nets were typically deployed in the late afternoon, soaked overnight, and retrieved the following morning. For each deployment, both the control and illuminated net were set.

Each boat captain then continued their fishing operations as per usual, fishing in the same areas for the same amount of time that they normally fished.

In order to detect statistical differences between the catch rates for the control illuminated nets, the mean CPUE values for both were compared using a t-test. Additionally, 2-sample t-tests were used to analyse differences in body size for sea turtles and target fish between control and illuminated nets. The study used the randomization test to analyse the catch data and test the null hypothesis that there would be no difference in sea turtle catch rate, total target catch rate, and CPUE between experimental and control nets. Data

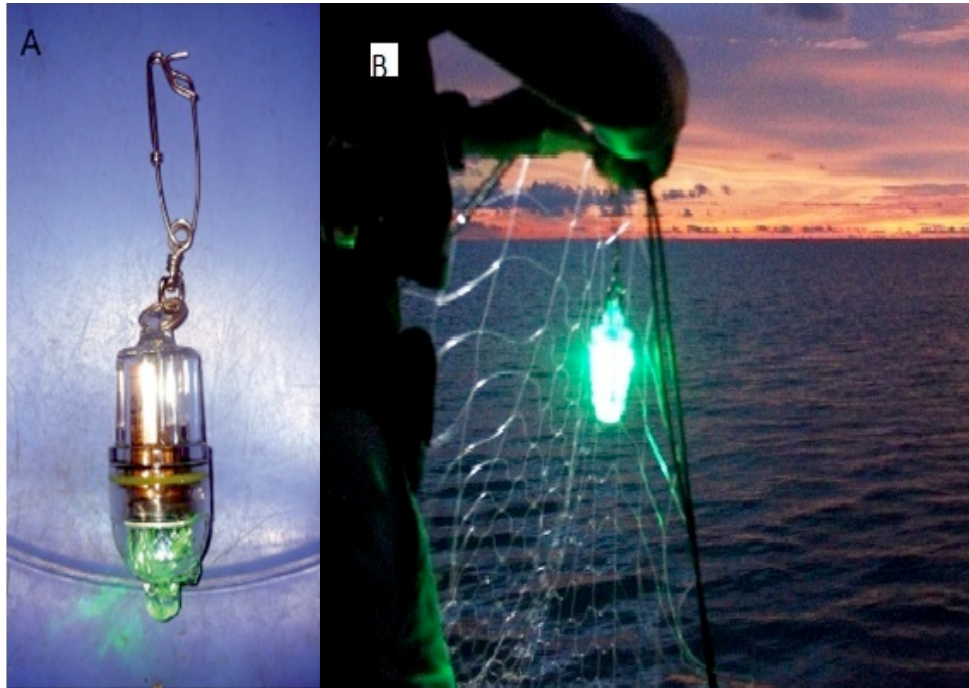


Figure 2. (A) Example of the LED light used during the study. (B) LED fitted on a bottom-set gillnet.

were re-sampled several times using the software RESAMPLING STATS for Excel (v. 4.0). This analysis measures the strength of evidence against a null hypothesis instead of estimating significance at a certain level.

Results

The study deployed 10 boats with paired control and experimental nets. The number of panels in each net varied slightly among boats and between trips as panels were sometimes added to increase target species catch, or were detached for repair. Therefore, net length varied, with control nets averaging $0.62 \text{ km} \pm 0.03 \text{ SE}$, while illuminated nets averaged $0.60 \text{ km} \pm 0.02 \text{ SE}$ (Table 1). Soak time for control nets averaged $17.10 \text{ hrs} \pm 0.39 \text{ SE}$, while experimental nets averaged $17.40 \text{ hrs} \pm 0.39 \text{ SE}$ (Table 1). Fishing effort was determined by combining net length and soak time ($\text{km} \times 24 \text{ hrs}$ soak). The mean fishing effort averaged $0.41 \pm 0.02 \text{ SE}$ ($\text{km} \times 24 \text{ h}$) for control nets, while illuminated

nets averaged $0.40 \pm 0.01 \text{ SE}$ ($\text{km} \times 24 \text{ h}$) (Table 1).

A total of 86 sea turtles were caught during the study period. Of these, 56 were caught in the control nets constituting 41 green, 9 hawksbills, 5 loggerheads, and 1 olive ridley turtle. The illuminated nets caught 30 turtles of which 21 were green, 5 hawksbills and 4 loggerhead turtles.

Analysis with the two-sample t-test indicated that sea turtle CPUE was significantly higher in control nets (mean CPUE = $1.40 (\pm 0.16 \text{ SE})$) as compared with experimental nets (mean CPUE = $0.50 \pm 0.06 \text{ SE}$), indicating a 64.3% reduction in mean catch rate ($p = 0.04$) (Table 3; Fig. 3). The paired nets were concurrently used to examine the effects of LED illumination on total target fish catch rates, composition and weight. A total of 12,987 individual target fish (46,581 kgs) were kept for market. Control nets caught 695 target fish (23,539 kgs) with a mean CPUE

Table 1. Summary of fishing effort by net type (control = without LED illumination, illuminated = with LED illumination) for paired gill net sets in the study area.

| Net type | Sets | Set duration (h) | | Net length (km) | | Fishing effort ($\text{km} \times 24 \text{ h}$) | |
|-------------|------|------------------|------------|-----------------|-----------|--|-----------|
| | | Mean \pm SE | Range | Mean \pm SE | Range | Mean \pm SE | Range |
| Control | 80 | 17.10 ± 0.39 | 2.83-24.07 | 0.62 ± 0.03 | 0.32-1.28 | 0.41 ± 0.02 | 0.07-1.10 |
| Illuminated | 80 | 17.40 ± 0.39 | 3.75-24.33 | 0.60 ± 0.02 | 0.32-1.15 | 0.40 ± 0.01 | 0.09-0.75 |

Table 2. Summary of target species and sea turtles (number caught) by net type (control = without LED illumination, illuminated = with LED illumination).

| Net type | Sets | Total effort (Km x 24 h) | Target species caught | Turtles caught |
|-------------|------|--------------------------|-----------------------|----------------|
| Control | 80 | 48.96 | 695 | 56 |
| Illuminated | 80 | 47.71 | 603 | 30 |

Table 3. The outputs and mean catch per unit effort of target species and sea turtles (control = net without LED illumination, illuminated = net with LED illumination).

| Response variable | Mean CPUE Control (mean \pm SE) | Mean CPUE Illuminated (mean \pm SE) | % diff. | p |
|-------------------|------------------------------------|--|---------|------|
| Target Species | 10.62 \pm 0.71 | 10.35 \pm 0.86 | -2.5 | 0.78 |
| Sea turtles | 1.40 \pm 0.16 | 0.50 \pm 0.06 | -64.28 | 0.04 |

of 10.62 \pm 0.71 SE, whereas experimental nets caught 603 target fish (23,042 kgs) with a mean CPUE of 10.35 \pm 0.86 SE (Table 3; Fig. 3), which was statistically similar ($P = 0.78$).

Fishermen interviewed after the completion of the project generally believed that the lights were effective at reducing marine turtle bycatch in their gillnets, although turtle bycatch was still present.

Discussion, conclusion and recommendations

Artisanal fishing in Kenya is a major source of income for more than 300,000 people in coastal communities with few economic opportunities other than those related to fishing (Fisheries Department, 2014). These important fisheries also account for significant sea turtle mortality (WWF, 2009). The purpose of this study was to investigate bycatch measures to reduce turtle bycatch without compromising target

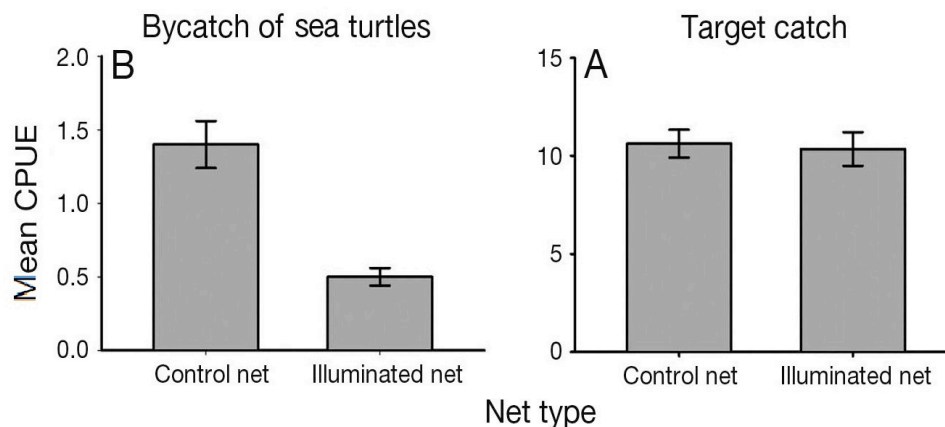


Figure 3. (A) Comparison of the mean CPUE of sea turtles between control and illuminated nets, and (B) comparison of the mean CPUE of target species between control (without LED illumination) and illuminated (with LED illumination) nets, showing no significant difference.

catches. This study showed that green LEDs attached to bottom-set gillnets in northern Kenya considerably reduce sea turtle bycatch, without adversely affecting target species catch rates. This technique could potentially serve as an effective sea turtle bycatch reduction device (BRD) for this type of fishery. This study demonstrates that, if this intervention was well managed and widely implemented, it could potentially promote the long-term stability of both sea turtle populations and local fisheries.

Sea turtles interact with gillnets globally (Wallace *et al.*, 2010). It will therefore be important to replicate this study in multiple locations and fisheries to assess the effectiveness of net illumination with a variety of gears, environmental conditions, and catch compositions (Southwood *et al.*, 2008; Gilman *et al.*, 2010). In order to effectively implement this BRD or other mitigation methods, any future studies need to consider costs and implications for fishermen, impacts on the catch of their target species, and the effect on other bycatch species (Cox *et al.*, 2007). Trials of this BRD in small-scale fisheries could serve as an important step in the global conservation of sea turtles. The cost of LEDs spread across multiple years still represents an untenable amount for Kenya's artisanal fishers. This means that efforts are needed at national or international levels to leverage financial support if this BRD is to be broadly implemented. To encourage this support, it would be useful to calculate the approximate cost (LEDs, gear etc) of preventing a single sea turtle interaction. This could then also be used to compare the costs of alternative conservation measures such as fisheries closures, time-area based closures, and development of marine reserves (Balmford *et al.*, 2004; McClanahan *et al.*, 2006).

Notwithstanding the challenges of implementing net illumination in artisanal fisheries (e.g. cost, LED design, fisher awareness), the results from this study emphasize the effectiveness of controlled fisheries experiments for the testing of bycatch reduction measures in artisanal gillnet fisheries. Future studies on net lighting should examine possible effectiveness as a multi-taxa technological tool for seabirds and marine mammals as these animals also rely on visual cues to a large extent (Jordan *et al.*, 2013; Martin and Crawford 2015). In addition, continued development of LED lights and their power sources could improve their efficiency and ensure optimal performance. Solar powered LEDs could also be developed in order to reduce the cost and waste associated with batteries.

Fishermen involved in the trials were mainly positive and provided essential feedback, which included the suggestion that LED light sticks should be designed specifically for net fisheries. Such continuous associations with fishermen and their communities will be critically important for the continued development and testing of net illumination as well as other bycatch reduction strategies for artisanal fisheries.

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